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Key Points:

- Positive δ²⁰²Hg shifts together with mercury concentration in lake sediment cores from pre-industrial to present-day period are a widespread phenomenon
- Magnitudes of δ^{202} Hg and Δ^{199} Hg changes are determined by the pre-industrial or baseline values in the sediment cores rather than the degree of mercury input
- Sediment δ²⁰²Hg can be used as a proxy for monitoring changes in anthropogenic mercury sources for the Minamata Convention on Mercury

Supporting Information:

Supporting Information may be found in the online version of this article.

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LEE ET AL.

Spatiotemporal Characterization of Mercury Isotope Baselines and Anthropogenic Influences in Lake Sediment Cores

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Abstract Increasing mercury isotope ratios from pre-industrial (1510–1850) to present-day (1990– 2014) in lake sediment cores have been suggested to be a global phenomenon. To assess factors leading to spatiotemporal changes, we compiled mercury concentration (THg) and mercury isotope ratios in 22 lake sediment cores located at various regions of the world. We find that the positive δ^{202} Hg shifts together with THg increases from pre-industrial to present-day are a widespread phenomenon. This is caused by increased contribution of mercury from local to regional anthropogenic mercury emission sources, which lead to higher sediment δ^{202} Hg (-1.07 ± 0.69‰, 1 SD) than pre-industrial sediments (-1.55 ± 0.96‰, 1 SD). The positive Δ^{199} Hg shifts were observed in 15 lake sediment cores, which have low pre-industrial Δ^{199} Hg ($-0.20 \pm 0.32\%$) compared to the sediment cores with near-zero to positive pre-industrial Δ^{199} Hg $(0.08 \pm 0.07\%)$. The magnitudes of δ^{202} Hg $(r^2 = 0.09)$ and Δ^{199} Hg $(r^2 = 0.20$, both p > 0.05) changes from pre-industrial to present-day did not correlate with the magnitude of THg changes. Instead, the magnitudes of δ^{202} Hg and Δ^{199} Hg changes decreased with increasing pre-industrial δ^{202} Hg and Δ^{199} Hg values, suggesting that the baseline mercury isotope ratios play a more important role in determining the magnitude of mercury isotope changes compared to the degree of THg input. We suggest that the spatiotemporal assessments of δ^{202} Hg in lake sediment cores can be used as an important proxy for monitoring changes in anthropogenic mercury sources for the Minamata Convention on Mercury.

1. Introduction

Mercury is a globally distributed atmospheric pollutant, which travels long distances in the form of gaseous elemental mercury (Hg⁰). Gaseous Hg⁰ is removed from the atmosphere via the foliar uptake (Demers et al., 2013; Zhou et al., 2021) or via oxidation to Hg²⁺, which is readily deposited to the biosphere following sorption to particles (Hg_p) and/or precipitation (Selin, 2009). Since industrialization, anthropogenic activities alone have increased mercury emission by a factor of approximately 5 (Streets et al., 2017). Increased mercury loading to aquatic ecosystems can enhance microbial production of monomethylmercury (MeHg) (Benoit et al., 2003; Lindberg et al., 2007), which is a bioaccumulative toxin in food webs (Mergler et al., 2007). Humans are primarily exposed to MeHg via the consumption of fishery products (Sunderland, 2007).

In 2017, the Minamata Convention on Mercury (MC), a multilateral agreement to mitigate anthropogenic mercury emissions and human health from mercury pollution, has entered into force (UNEP, 2019). As a part of the MC, provisions have been established for a global monitoring program and a convention effectiveness evaluation (Article 19, 22) to understand spatiotemporal changes in mercury levels as well as its sources, processes, and fate in various environmental media. Since then, numerous studies have assessed temporal trends of mercury in diverse atmospheric samples (Hg⁰, Hg²⁺, Hg_p, precipitation) (Cheng et al., 2017; Dommergue et al., 2016) and biota (fish, bird eggs, polar bear) (Blukacz-Richards et al., 2017; Lee et al., 2016; McKinney et al., 2017) to gather insights on the changes in emissions, deposition, and ecosystem fate of mercury. Natural archives of sediment, peat, and ice cores have also been used to quantify long-term changes in the deposition of various atmospheric mercury species (Engstrom et al., 2014; Enrico et al., 2017; Zdanowicz et al., 2016; Zerkle et al., 2020). In particular, mercury in sediment cores of lakes with large surface area relative to watershed area is thought to reflect delivery via precipitation and is less prone to diagenetic and hydrodynamic effects resulting in mobility and release compared to mercury in peat and marine sediment cores (Asmund & Nielsen, 2000; Cooke et al., 2020; Wasik et al., 2015). Recent studies reported consistent 3–5 fold enrichments in anthropogenic mercury relative to the pre-industrial period in lake sediment cores collected from various remote regions (Amos et al., 2015; Engstrom et al., 2014; Fitzgerald et al., 2005).

While mercury concentrations in lake sediment cores have been evaluated on a broad spatiotemporal scale, sources as well as atmospheric and biogeochemical processes leading to spatiotemporal trends in mercury isotope ratios are only beginning to be understood. Mercury isotopes undergo mass-dependent fractionation (MDF; δ^{202} Hg) via various biotic and abiotic processes including methylation, demethylation, and mineral sorption (e.g., Blum et al., 2014; Kwon et al., 2020). Significant mass-independent fractionation of odd-mass number isotopes (MIF_{odd}; Δ^{199} Hg, Δ^{201} Hg) has been observed primarily during photo-reduction and photo-degradation of Hg²⁺ and MeHg, respectively (Bergquist & Blum, 2007). Significant MIF of even-mass number isotopes (MIF_{even}; Δ^{200} Hg) is thought to occur via photo-oxidation of Hg⁰ in the atmosphere (Chen et al., 2012; Cai & Chen, 2016). Given the limited processes resulting in MIF, Δ^{199} Hg has been used to interpret changes in mercury sources and/or processes in natural archives including sediment and ice cores (Guédron et al., 2016; Kurz et al., 2019; Yin, Feng, et al., 2016; Zerkle et al., 2020).

Recently, there has been a consensus from the literature that increasing Δ^{199} Hg (by 0.1–0.3‰) in lake sediment cores from pre-industrial to present-day period is a global phenomenon (Guédron et al., 2016; Kurz et al., 2019; Yin, Feng, et al., 2016). To explain this phenomenon, multiple hypotheses have been proposed. Kurz et al. (2019) proposed that the changes in global atmospheric chemical composition caused by increased levels of anthropogenic atmospheric pollutants can enhance Hg²⁺ photo-reduction in precipitation and result in positive Δ^{199} Hg shifts in a remote Wyoming lake, U.S. In remote Tibetan lakes, decreased ice cover resulting in increased Hg²⁺ photo-reduction in the water column was suggested as the main driver for the positive Δ^{199} Hg shifts (Yin, Feng, et al., 2016). In contrast, Guérdron et al. (2016) measured Δ^{199} Hg in a French lake surrounded by multiple mercury point sources and suggested that increased industrial mercury sources, characterized by less negative Δ^{199} Hg compared to geogenic sources, is responsible for the Δ^{199} Hg changes. In addition, a recent study by Lepak et al. (2020) observed positive temporal shifts in Δ^{199} Hg and δ^{202} Hg in nine lake sediment cores collected from remote regions of North America. The authors suggested that the magnitude of changes in mercury isotope ratios from pre-industrial to present-day period is governed by the amount of atmospheric mercury loading. As such, it is unclear whether the positive Δ^{199} Hg and δ^{202} Hg shifts are driven by site-specific or global changes in mercury sources or processes. The understanding of the linkages between the magnitude of mercury input and changes in mercury isotope ratios across a large spatiotemporal scale will also be important for using lake sediment cores as a proxy for global monitoring and effectiveness evaluation of the MC.

Here, we compiled mercury concentration and mercury isotope ratios of lake sediment cores from various locations of the world. Our aim was to (a) characterize spatiotemporal differences in mercury levels and isotope ratios, (b) gather insights into source- and/or process-driven changes in mercury isotope ratios, and (c) identify factors governing the magnitude of changes in mercury isotope ratios relative to the changes in mercury input. We discuss our results in relation to the global monitoring and effectiveness evaluation objectives of the MC.

2. Materials and Methods

Total mercury (THg) concentration and mercury isotope ratios (δ^{202} Hg, Δ^{199} Hg, Δ^{200} Hg) of 22 lake sediment cores were compiled from the literature (Cooke et al., 2013; Gray et al., 2013; Guédron et al., 2016; Kurz et al., 2019; Lepak & Janssen, 2020; Ma et al., 2013; Yin et al., 2016a, 2016b). For compilation, we first screened for sediment cores, which reported both THg and mercury isotope ratios prior to significant local anthropogenic mercury influences and/or those dating back to pre-industrialization. We then screened for sediment cores, which followed the standard mercury isotope ratio reporting and measurement system (Bergquist & Blum, 2009; Blum & Johnson, 2017) of using NIST SRM 3133 and NIST RM 8610 (also known as UM-Almadèn) as a bracketing and analytical standard, respectively, to ensure data quality. All lake sediment cores compiled in this study were dated by ²¹⁰Pb using isotope-dilution, alpha spectrometry methods, and the constant rate of supply dating model (Table S1). The complete list of THg concentration and mercury isotope ratios associated with each sediment core and its depth and age dating are reported in our compiled data.

Based on the inventories of all-time anthropogenic mercury emissions from Streets et al. (2017), we set the pre-industrial period as years between 1510 and 1850. The present-day is set to years between 1990 and 2014, which is consistent with those referenced by Amos et al. (2015) and Li et al. (2020) using empirical assessments of mercury in various natural archives. Note that year 2014 is the earliest mercury isotope measurement available in our compiled data. To characterize mercury levels and isotope baseline of the pre-industrial period, we calculate the average THg concentration and mercury isotope ratios spanning the years between 1510 and 1850 for each sediment core. For sediment cores without data prior to 1850 (5 cores; Phantom, Cleaver, McLurg Lake in Manitoba, Canada, Lake Luitel in France, Lake Michigan in the U.S.) (Guédron et al., 2016; Ma et al., 2013; Yin, Lepak, et al., 2016), we use the average THg concentration and mercury isotope ratios of three data points representing the deepest layer, which span years between 1873 and 1913. While these years represent the time during industrialization, the THg and mercury isotope ratios reflect those prior to significant local anthropogenic mercury emissions as suggested by the individual authors.

The spatiotemporal changes in mercury levels and isotope ratios are evaluated by (a) comparing the average THg concentration and mercury isotope ratios between the pre-industrial and present-day period, (b) by testing the significance of differences in THg concentration and mercury isotope ratios between the pre-industrial and present-day period, and (c) by assessing the significance of temporal trends in mercury isotope ratios for each sediment core. We calculate the average THg concentration and mercury isotope ratios reflecting years between 1990 and 2014 (present-day) for each sediment core. Several cores had only one data point spanning these years (three cores; El Junco, Laguna Negrilla in Peru, Perfect Lake in the Arctic). We did not take the average of the most recent three data points because they span years between 1964 and 1983, which are too old to include. We set the significant difference in the average mercury isotope ratios between the pre-industrial and present-day period as changes greater than the analytical uncertainty (2SD) reported by individual studies compiled here; $\pm 0.09\%$ for δ^{202} Hg, $\pm 0.06\%$ for Δ^{199} Hg, and $\pm 0.05\%$ for Δ^{200} Hg (Table S2). The Kruskal–Wallis test, which assumes nonparametric data, was also used to test the significance of differences in THg and mercury isotope ratios between the pre-industrial and present-day periods. Based on the results of Shapiro-Wilk test for normality, the THg concentrations and mercury isotope ratios, except for the present-day unremote Δ^{200} Hg, were not normally distributed (Table S3, Figure S1), which confirm the use of the Kruskal-Wallis test. The H-value is the test statistic for the Kruskal-Wallis test to calculate the *p*-value and the *p*-value (<0.05) indicates the significance of correlation between two variables. The two-tailed conventional linear trend test was used to verify the significance of temporal trends in THg and mercury isotope ratios over the pre-industrial to present-day period for each core. The slope indicates the direction of trend between two variables and the p-value (<0.05) indicates the significance of temporal trend (Table S4).

To characterize spatial differences in the changes in mercury levels and isotope ratios, the magnitude of differences in THg concentration and mercury isotope ratios between the pre-industrial and present-day periods is evaluated across different sampling locations of the lake sediment cores and between remote and unremote regions. The remote and unremote regions are categorized based on the proximity to local anthropogenic mercury sources and those characterized by the individual authors. Specifically, one sediment core from Lake Ballinger, Washington, U.S. (Gray et al., 2013) and three sediment cores from Manitoba, Canada (Phantom, Cleaver, McLurg Lake; Ma et al., 2013) are known to be impacted by local point mercury sources of lead, copper, and zinc smelters. One sediment core from Lake Luitel, Grenoble, France (Guérdron et al., 2016) and three sediment cores from Lake Michigan, U.S. (Yin, Lepak, et al., 2016) have received diverse anthropogenic mercury sources of manufacturing industries and mining activities, originated from rapid industrialization since the mid-1850s. Five sediment cores collected from the national forest in Wyoming, U.S. (Lost Lake; Kurz et al., 2019), the Peruvian Andes and Galápagos archipelago, Ecuador (El Junco, Laguna Negrilla; Cooke et al., 2013), and Tibetan Plateau, China (Qinghai, Nam Co; Yin, Feng,

et al., 2016) are thought to be one of the most remote locations in the world. The remaining nine sediment cores are from the study by Lepak et al. (2020), in which the authors described the lakes as generally remote and undisturbed from Alaska and Minnesota, U.S. and Newfoundland, Canada.

3. Results and Discussion

3.1. Spatiotemporal Differences in Mercury Levels and Isotope Ratios

The locations of the sediment cores are skewed to the Northern Hemisphere and specifically to North America (Figure 1). Elevated THg concentrations were observed in Laguna Negrilla, Peru (228–684 ng/g) (Cooke et al., 2013) and Phantom Lake, Manitoba, Canada (496–941 ng/g) (Ma et al., 2013) during the pre-industrial period (1510–1850) (Figure 1a), which were suggested to be impacted by mercury used during historical mining. Without these lake sediments (Laguna Negrilla, Phantom Lake), the THg concentrations were much lower (average; 99 ± 57 ng/g, range; 17–249 ng/g, n = 142) and similar to those of pre-industrial sediments reported previously (approximately 70 ng/g) (Fitzgerald et al., 2005; Engstrom et al., 2007, 2014). The present-day sediments had significantly higher and wide ranges in THg concentration (median; 265 ± 2933 ng/g, range; 38–19,076 ng/g, n = 103) compared to the pre-industrial sediments (Kruskal–Wallis test, p < 0.05, Table S5).

To understand the spatial difference in the magnitude of THg concentration change, we compared the magnitude of THg change between the pre-industrial and present-day period and observed a wide range across the sites (0.53- to 25-fold increase; Figure 1a). For present-day samples, significantly higher THg concentrations were observed in the unremote lakes (median; 620 ± 4140 ng/g, range; 127-19,076 ng/g, n = 47) compared to the remote sediments (median; 192 ± 104 ng/g, range; 38-497 ng/g, n = 56, Kruskal-Wallis test, p < 0.05, Table S5). The present-day unremote sediments showed 2.9- to 25-fold increases in THg relative to their respective pre-industrial sediments. The present-day remote sediments, without Laguna Negrilla, Peru, showed 1.3- to 5.1-fold increases in THg, consistent with those reported from other lake sediment cores collected from remote regions (3-5 fold) (Amos et al., 2015; Engstrom et al., 2014; Fitzgerald et al., 2005, Li et al., 2020). Among the remote sediments, a reduction in THg (0.53-fold) was observed in Laguna Negrilla, Peru (Figure 1a), consistent with the cessation of widespread historical mercury mining for pigment and precious metal extraction (1532-1821 AD) (Cooke et al., 2013). Without Laguna Negrilla, Peru, relatively small increases in THg (1.1- to 2.1-fold increase) were observed in the Tibetan Plateau, China, Alaska, U.S., and the Peruvian Andes, Educador, which are some of the least urbanized and industrialized locations in the world. Higher increases in THg (2.0- to 5.1-fold) and those comparable with the unremote sediments were detected from the Western (Wyoming) and Midwest, U.S. (Wisconsin, Minnesota) and Newfoundland, Canada, suggesting that these locations may be influenced by the transport of regional anthropogenic mercury sources. Our results suggest that the spatial difference in the magnitude of THg increase is governed by the proximity to local and regional anthropogenic mercury sources.

Consistent with the changes in THg, the present-day sediments showed statistically higher mercury isotope ratios (δ^{202} Hg, Δ^{199} Hg, Δ^{200} Hg) compared to the pre-industrial sediments (Kruskal–Wallis test, all p < 0.05, Table S5, Figures 1 and 2). The average δ^{202} Hg and Δ^{199} Hg of the pre-industrial sediments were $-1.55 \pm 0.96\%$ (-5.04 to 0.37%, n = 138) and $-0.15 \pm 0.28\%$ (-0.79 to 0.29%, n = 138), respectively. Only 17 out of 22 cores reported Δ^{200} Hg (Figure S2), which showed an average value of 0.03 \pm 0.07‰ (-0.09 to 0.23‰, n = 122). The average δ^{202} Hg, Δ^{199} Hg, and Δ^{200} Hg of the present-day sediments were -1.07 ± 0.69 ‰ $(-3.68 \text{ to } -0.16\%, n = 100), 0.07 \pm 0.22\%$ $(-0.44 \text{ to } 0.48\%, n = 100), and 0.06 \pm 0.06\%$ (-0.03 to 0.20%, n = 100)n = 78), respectively. The magnitude of changes in mercury isotope ratios varied widely across geographic locations (Figures 1b and 1c). Even within a single lake, Lake Michigan, the magnitude of changes in mercury isotope ratios from the pre-industrial to present-day period was highly variable. This suggests that, unlike THg, the magnitude of changes in mercury isotope ratios are not determined by the proximity to anthropogenic mercury sources, which we discuss further in Section 3.3. Among the present-day sediments, the remote sediments displayed statistically higher Δ^{200} Hg (Kruskal–Wallis test, p < 0.05, Table S5), but not δ^{202} Hg (p = 0.77) and Δ^{199} Hg (p = 0.06), compared to the unremote sediments. The average Δ^{200} Hg of the remote and unremote sediments were $0.09 \pm 0.07\%$ (n = 46) and $0.02 \pm 0.03\%$ (n = 32), respectively (Figure 2c). The unremote sediments were characterized by a narrow range in δ^{202} Hg ($-0.93 \pm 0.31\%$, n = 52) and were statistically similar with the remote sediments, which had a wider range in δ^{202} Hg (-1.21 ± 0.93%),



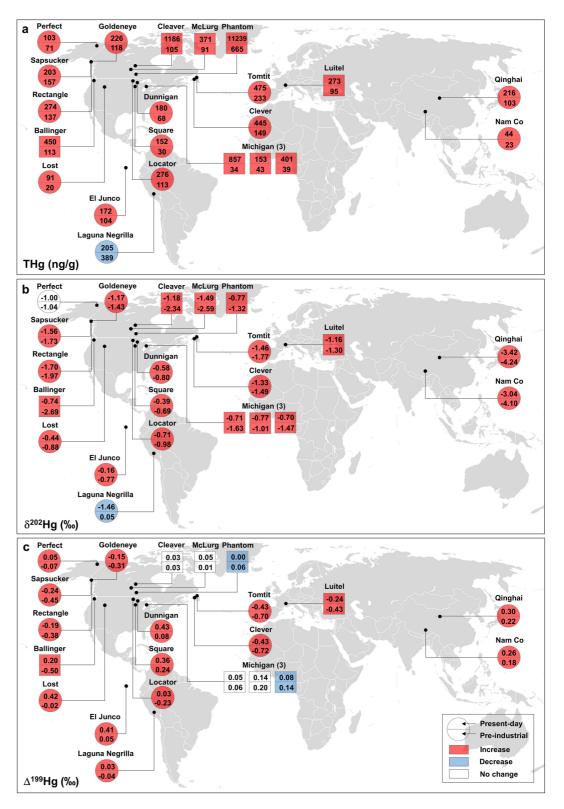


Figure 1. Spatiotemporal changes in (a) THg concentration (ng/g), (b) δ^{202} Hg (‰), and (c) Δ^{199} Hg (‰) from the preindustrial to present-day period in all lake sediment cores. The sediment cores shown in circles represent remote sites and the sediment cores in squares represent unremote sites.



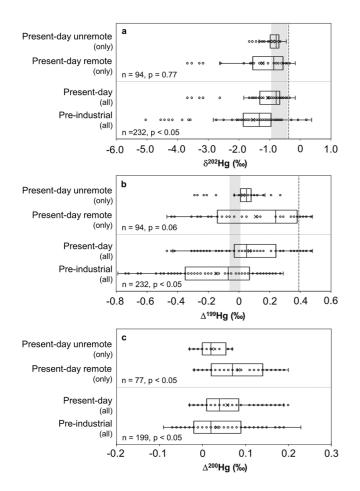


Figure 2. The ranges of (a) δ^{202} Hg (‰), (b) Δ^{199} Hg (‰), and (c) Δ^{200} Hg (‰) of the pre-industrial and present-day sediments. The present-day sediments are further divided into remote and unremote. The shaded areas represent the ranges of modern-day δ^{202} Hg (-0.94 to -0.44‰) and Δ^{199} Hg (-0.07 to 0.01‰) of anthropogenic THg emission source end member estimated by Sun et al. (2016). The dotted lines represent the median δ^{202} Hg (-0.38‰) and Δ^{199} Hg (0.39‰) of precipitation compiled by Kwon et al. (2020).

n = 48, Figure 2a). Both the unremote $(0.03 \pm 0.11\%, n = 52)$ and remote sediments $(0.12 \pm 0.29\%, n = 48)$ showed near-zero Δ^{199} Hg (Figure 2b).

The mercury isotope ratios of the pre-industrial sediments appear to reflect both natural and anthropogenic mercury delivered via atmospheric deposition and runoff from the adjacent watershed. Previous evaluations of anthropogenic mercury enrichments in sediment and peat cores have suggested that historical anthropogenic activities (since 1450 AD) such as Spanish colonial mercury and silver mining (Amos et al., 2015) as well as deforestation and biomass combustion (Li et al., 2020) have liberated substantial amounts of mercury into the atmosphere. The resultant Hg⁰ would have accumulated within foliage prior to watershed runoff or deposited to lake sediments upon atmospheric oxidation. Sun et al. (2016), who estimated the isotopic end members for anthropogenic mercury emission sources, suggested that historical anthropogenic mercury emissions (approximately 1850s), originated via mining and liquid mercury use in gold and silver refining processes, have ranges of median δ^{202} Hg of -1.19 to -1.10% (THg, Hg⁰, Hg²⁺, Hg_p), which are similar to the pre-industrial sediment δ^{202} Hg compiled here but much lower than those estimated for modern-day anthropogenic mercury emission sources. The estimated Δ^{199} Hg are similar between the historical (-0.05 to -0.02%) and modern-day anthropogenic mercury emission sources (-0.06 to -0.02%); Sun et al., 2016). The atmospheric deposition of both natural and historical anthropogenic mercury emission sources, together with watershed runoff of Hg⁰ sequestered in foliage, which is known to cause measurable negative δ^{202} Hg shifts in foliage relative to atmospheric Hg⁰ (Demers et al., 2013), would explain the low δ^{202} Hg in the pre-industrial sediments. As for Δ^{199} Hg, the influence of terrestrial mercury sources (soil, litterfall), characterized by negative Δ^{199} Hg (-0.49 to -0.14‰; Demers et al., 2013; Jiskra et al., 2015, 2017), also explain the low Δ^{199} Hg in the pre-industrial sediments.

The significantly higher THg and mercury isotope ratios observed in the present-day sediments compared to the pre-industrial sediments can be explained by the influence of modern-day anthropogenic mercury emission sources. Anthropogenic emissions of Hg²⁺ and Hg_p, which readily deposit to lakes relative to Hg⁰, have increased in the modern-day period owing to the activities in by-product sectors (i.e., fossil fuel combustion) (Sun et al., 2016). Sun et al. (2016) estimated significantly higher δ^{202} Hg

in the modern-day (approximately 2010s) anthropogenic THg emission sources (median δ^{202} Hg; -0.70%, range; -0.94 to -0.44%) relative to the pre-industrial anthropogenic emissions and the average pre-industrial sediment δ^{202} Hg compiled in this study. The median δ^{202} Hg (-0.84‰) reflecting modern-day anthropogenic Hg^{2+} and Hg_p emissions also showed similar estimated values relative to THg (Sun et al., 2016). The anthropogenic mercury emission source end members estimated by Sun et al. (2016) are relatively similar with the average δ^{202} Hg and Δ^{199} Hg of the present-day unremote sediments impacted by anthropogenic activities and those that displayed narrow ranges in δ^{202} Hg and Δ^{199} Hg (Figures 2a and 2b). Precipitation samples influenced by varying degrees of anthropogenic mercury emission sources and those recently compiled by Kwon et al. (2020) also show higher δ^{202} Hg (median -0.38%; -4.27 to 0.31%, n = 122) and particularly high Δ^{199} Hg (median 0.39%; -0.87 to 1.57%, n = 122) compared to the average pre-industrial sediments. Given the approximately 4-fold increase in the global atmospheric mercury deposition since industrialization (Amos et al., 2015), it is likely that the widespread anthropogenic activities have increased the amount of mercury available for deposition to lake sediments even at remote sites. This explains the absence of significant differences in δ^{202} Hg and Δ^{199} Hg between the present-day remote and unremote sediments. Moreover, even with the modern-day increase in Hg⁰ available for foliar uptake and runoff to lake sediments, the estimated higher δ^{202} Hg of Hg⁰ in the modern-day anthropogenic mercury emission sources (median: -0.56%) relative to historical anthropogenic mercury emission sources (median: -1.12%) (Sun et al., 2016) may explain the overall δ^{202} Hg increase in the present-day sediment.

In regards to Δ^{200} Hg, a recent study by Lepak et al. (2020) suggested that the spatial variability in Δ^{200} Hg among remote lake sediments is driven by the relative contribution of atmospheric (precipitation Hg²⁺) versus watershed (foliage Hg⁰) mercury sources (Lepak et al., 2020). While this may be the case for our present-day remote sediments, the significantly higher Δ^{200} Hg observed in the present-day remote sediments relative to the unremote sediments (Figure 2c) may be explained by a higher extent of mercury photo-oxidation prior to deposition. The lakes designated as remote in this study are located far from local anthropogenic mercury emissions sources, as suggested by the individual authors, and receive mercury that has been subjected to regional transport. The Δ^{200} Hg has been used as a conservative tracer to evaluate the relative importance of precipitation derived mercury input to surface reservoirs and/or the extent of atmospheric mercury photo-oxidation (Blum & Johnson, 2017). Here, we compiled wet mercury deposition fluxes from Lepak and Janssen (2020), who reported measured fluxes to nine remote lakes in North America and those compiled here, and from Mulvaney et al. (2020), who simulated a global atmospheric chemistry transport model for mercury (GEOS-Chem; Table S6). There was no significant difference in wet mercury deposition fluxes between the remote and unremote lakes (Kruskal–Wallis test, p < 0.05), indicating that the extent of precipitation derived mercury input is not responsible for the Δ^{200} Hg difference between the remote and unremote sediments. The watershed area to lake area ratio, an indicator for the relative input of atmospheric (Hg²⁺) versus watershed mercury (Hg⁰) input, also showed no clear differences between the remote and unremote lakes (Table S6). We speculate that a higher extent of Hg⁰ photo-oxidation, resulting in a higher Δ^{200} Hg and a lower δ^{202} Hg in Hg²⁺ compared to Hg⁰ (Cai & Chen, 2016), during regional transport prior to deposition may explain the significantly higher Δ^{200} Hg and slightly lower δ^{202} Hg observed in the remote sediments compared to the unremote sediments.

As for the present-day remote sediments, the relative contribution of atmospheric versus watershed mercury sources appears to explain the spatial variability in Δ^{200} Hg. We observed significantly higher δ^{202} Hg $(-0.54 \pm 0.16\%)$, Δ^{199} Hg $(0.28 \pm 0.19\%)$, and Δ^{200} Hg $(0.13 \pm 0.05\%)$ in five lakes (Dunnigan and Square, Minnesota, Locator, Wisconsin, Lost, Wyoming in the U.S., El Junco, Peru; Figures 1b, 1c, S2), characterized by high wet mercury deposition fluxes and small watershed area to lake area ratios, compared to other remote lakes compiled in this study (δ^{202} Hg; -1.94 ± 0.87%, Δ^{199} Hg; -0.07 ± 0.28%, Δ^{200} Hg; 0.03 ± 0.03%, Kruskal-Wallis test, all p < 0.05, Table S5). Lepak et al. (2020) reported higher wet mercury deposition fluxes (7.7–9.4 µg/m²/yr) coupled with small watershed area to lake area ratios (2.0–8.2) in Dunnigan, Locator, and Square compared to other remote lakes in North America (wet mercury deposition fluxes; $1.5-4.6 \,\mu g/m^2/$ yr, watershed area to lake area ratios; 1.2 to 13). The present-day wet mercury deposition fluxes simulated by GEOS-Chem and those reported by Mulvaney et al. (2020) also show much higher values (8.2–12 μ g/m²/ yr) and small watershed area to lake area ratios (2.1-2.2) in El Junco and Lost compared to other remote lakes (Table S6). All in all, our compilation study suggests that mercury isotope ratios in lake sediment cores can be used to evaluate the large-scale influence of modern-day anthropogenic mercury emission sources as well as the relative importance of mercury input pathways (atmospheric vs. watershed runoff) across remote regions. While further study is required to verify processes affecting mercury isotope ratios in the atmosphere prior to deposition to lake sediments, we suggest that sediment Δ^{200} Hg may be used to gain insights into the influence of local versus regional anthropogenic mercury deposition.

3.2. Sources and Processes Governing Temporal Changes

Results from the conventional linear trend test show widespread increasing trends of THg and mercury isotope ratios in the lake sediment cores from the pre-industrial to present-day period. Most of the sediment cores showed a statistically significant increasing linear trend in THg with the 95% confidence level (*p* value ≤ 0.05 ; Table S4) while one lake sediment shows a significant decreasing trend in THg (Laguna Negrilla). Based on the analytical uncertainty estimates, significant positive shifts in δ^{202} Hg (>2SD of 0.09‰) were observed in 20 sediment cores from the pre-industrial to present-day period, except for Perfect Lake in the Arctic (Lepak et al., 2020) and Laguna Negrilla, Peru (Cooke et al., 2013) (Figure 1b). Sixteen and three lake sediment cores showed a significant positive shift in Δ^{199} Hg (>2SD of 0.06‰) and Δ^{200} Hg (>2SD of 0.05‰), respectively (Figures 1c and S2). According to the linear trend test, 17 out of 22 lake sediment cores showed



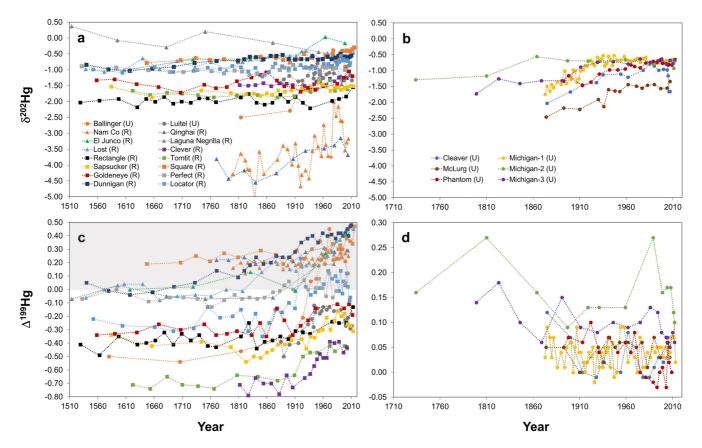


Figure 3. Temporal trends in δ^{202} Hg (a), (b) and Δ^{199} Hg (c), (d) of all lake sediment cores. The lake sediment cores designated as U and R indicate unremote and remote, respectively. The shaded area in panel c represents sediment cores (Nam Co, Qinghai in Tibet, Dunnigan, Square in Minnesota, U.S.), which exhibited positive Δ^{199} Hg shifts even with elevated Δ^{199} Hg baselines. Panel d represents sediment cores that showed no significant shifts in Δ^{199} Hg (Cleaver, McLurg in Manitoba, Canada, two locations in Lake Michigan) and sediment cores that had significant negative Δ^{199} Hg shifts (Phantom in Manitoba, Canada, one location in Lake Michigan).

a statistically significant increasing trend in δ^{202} Hg at the 95% confidence level (p value ≤ 0.05 ; Table S4). Fifteen lake sediment cores showed significant positive temporal trends in Δ^{199} Hg at the 95% confidence level and two cores showed significant negative trends in Δ^{199} Hg (Table S4). Overall, our results suggest that, in addition to Δ^{199} Hg, the positive shifts in THg concentration together with δ^{202} Hg are widespread phenomena across the lake sediment cores.

3.2.1. δ²⁰²Hg

The widespread positive δ^{202} Hg shifts from the pre-industrial to present-day period appear to be caused by increased contribution of modern-day anthropogenic mercury sources, which are characterized by a higher δ^{202} Hg compared to the pre-industrial sediments. Among the sediment cores compiled in this study, we observed significant negative relationships between δ^{202} Hg and 1/THg (p < 0.05, $r^2 = 0.21$ to 0.93), except for three lakes in North America (Goldeneye and Rectangle, Alaska, Perfect Lake in the Arctic) and Lake Qinghai, Tibet (Yin, Feng, et al., 2016) (p > 0.05, $r^2 = 0.03$ to 0.15; Table S7). This suggests that the majority of sediment cores depict an increasing contribution from anthropogenic mercury sources to the pre-industrial sediments, which have low THg and negative δ^{202} Hg (Figure 2).

The significant temporal δ^{202} Hg trends observed in the lake sediment cores further suggest that local to regional anthropogenic mercury sources, which share a similar δ^{202} Hg, contribute to the positive δ^{202} Hg shifts. As illustrated in Figure 3, both the remote and unremote lake sediment cores exhibited increasing trends in δ^{202} Hg with temporally corresponding peaks in THg and δ^{202} Hg (Figures 3a, 3b, and S3). The median δ^{202} Hg corresponding to the THg peaks is $-0.80 \pm 0.92\%$, which is within the ranges of precipitation impacted by anthropogenic activities and anthropogenic mercury emission end member estimated by

Kwon et al. (2020) and Sun et al. (2016), respectively. An exception was observed in Laguna Negrilla, which had a higher pre-industrial δ^{202} Hg compared to the anthropogenic mercury source end member and exhibited a decreasing temporal trend in THg and δ^{202} Hg, consistent with the cessation of widespread historical mercury mining for pigment and precious metal extraction (Cooke et al., 2013). The measurable reductions in THg concentrations relative to small changes in δ^{202} Hg in the recent decades (Figures 3a, 3b, and S3) may be explained by the absence of significant alterations in mercury sources. Amos et al. (2015), who compiled mercury concentrations in natural archives, also reported peaks in THg during the second half of the 20th century and declines in the recent decades. The authors attributed this to large-scale reductions in anthropogenic mercury emissions leading to rapid reductions in atmospheric mercury deposition and slow reductions in terrestrial reservoirs, which both supply mercury to lake sediments. Similarly, while the rapid and slow reductions in atmospheric deposition and watershed runoff of mercury may explain the overall declining trends in sediment THg concentrations, the input of constant mercury sources may explain minor changes in δ^{202} Hg.

Based on the similar δ^{202} Hg value between the median anthropogenic mercury emission source end member estimated by Sun et al. (2016) and the average δ^{202} Hg of the present-day unremote sediments, we speculate that processes affecting mercury in the atmosphere and within a lake system play minor roles in modifying δ^{202} Hg of anthropogenic mercury sources. While many processes are known to result in significant MDF (Blum et al., 2014; Kwon et al., 2020), previous studies have suggested that lake sediment cores are less prone to diagenetic and hydrodynamic effects resulting in mercury mobility and release compared to other natural archives (i.e., peat, marine sediment cores) (Asmund & Nielsen, 2000; Cooke et al., 2020; Wasik et al., 2015). This is supported by the studies that showed consistency in the levels and temporal trends in mercury in lake sediment cores, which have been collected at the same location in different periods (Percival & Outridge, 2013; Rydberg et al., 2008). A recent modeling study of mercury isotopes within a lake system also simulated small to negligible δ^{202} Hg changes in mercury input sources in the sediment during water column and diagenetic processes (Bessinger, 2014). From a temporal perspective, if changes in atmospheric and water column processes were significant enough to cause positive temporal trends in δ^{202} Hg, THg concentrations in the sediments would decrease, which is not what we observed. Experiments have shown that photo-reduction of Hg²⁺ and evaporation of Hg⁰ result in a higher δ^{202} Hg in the remaining solution compared to the product (Bergquist & Blum, 2007; Estrade et al., 2009; Ghosh et al., 2013). Increased Hg²⁺ photo-reduction in precipitation or in the water column would reduce the amount of mercury available for deposition to sediments. While increased mercury sorption to particles either in the atmosphere or in the water column may increase sediment THg, an experimental study by Wiederhold et al. (2010) reported a lower δ^{202} Hg in mercury bound to particles compared to those unbound in solution. We suggest that the increased contribution of modern-day anthropogenic mercury sources to the pre-industrial sediments explains the widespread δ^{202} Hg and THg shifts rather than temporal changes in processes affecting mercury in the atmosphere or within a lake system.

3.2.2. Δ^{199} Hg

In contrast to the previous argument that the positive Δ^{199} Hg shifts from pre-industrial to present-day period are a global phenomenon, the significant positive shifts were observed in a subset of lake sediment cores compiled in this study. Kurz et al. (2019) provided three hypotheses to explain the positive Δ^{199} Hg shifts observed in a remote lake sediment core from Wyoming, U.S.; (a) changes in the global atmospheric chemical composition, (b) changes in local processes including increased Hg²⁺ photo-reduction in the water column, and (c) increased anthropogenic mercury source contribution to lake sediments, which have more positive Δ^{199} Hg compared to geogenic sources. We propose a modified scenario from the latter in which the increased contribution of anthropogenic mercury sources is responsible for the positive Δ^{199} Hg shifts and the significant positive shifts are detected primarily in lake sediments with low pre-industrial Δ^{199} Hg.

We observed significant increasing temporal trends in Δ^{199} Hg in all remote (n = 13) and some unremote lake sediment cores (n = 2) (Figure 3c). Most of these lake sediment cores were characterized by negative pre-industrial Δ^{199} Hg (average -0.38%, -0.72 to -0.02%, n = 10) relative to the anthropogenic mercury emission source end member (Sun et al., 2016) and precipitation (Kwon et al., 2020), which have near-zero and positive Δ^{199} Hg, respectively. Seven lake sediment cores, characterized as near-zero and small positive pre-industrial Δ^{199} Hg (average 0.08%, 0.01-0.20%), showed no significant and significant reduction in temporal trends in Δ^{199} Hg, respectively (Figure 3d). It is possible that the higher extent of Hg²⁺ photo-reduction within the water column, resulting in a higher Δ^{199} Hg in the remaining mercury relative to Hg⁰ (Bergquist & Blum, 2007), may explain the small positive Δ^{199} Hg in these sediments. Alternatively, the pre-industrial unremote sediments with near-zero Δ^{199} Hg reflect local anthropogenic mercury sources. We note, however, that the pre-industrial unremote sediments had similarly low THg concentration (average 187 ng/g, 30–941 ng/g) compared to the remote sediments (average 147 ng/g, 18–684 ng/g, Figure 1a). While it is difficult to identify sources and/or processes resulting in more elevated Δ^{199} Hg in certain pre-industrial unremote sediments relative to others, the mixing between two similar end members (near-zero to positive pre-industrial vs. anthropogenic Δ^{199} Hg) appears to be responsible for the absence of significant temporal Δ^{199} Hg trends in certain sediment cores.

While only a few processes are known to cause significant MIF (Blum et al., 2014; Kwon et al., 2020), large global environmental changes may have contributed partly to the temporal shifts in Δ^{199} Hg. Previous studies have collectively suggested that, on a global basis, the amount of precipitation has increased significantly since the early 1900s (Ren et al., 2013). This observation is, however, in contrast with the absence of widespread and significant temporal modifications in Δ^{200} Hg (>2SD of 0.05‰). On a local scale, the effects of gradual and dramatic changes such as the ice cover and erosional and land-use change may lead to positive temporal shifts in Δ^{199} Hg. In fact, several exceptions were observed in Tibetan lakes (Qinghai, Nam Co), and Dunnigan and Square in Minnesota, which showed increasing temporal Δ^{199} Hg trends even with positive pre-industrial Δ^{199} Hg (average; 0.15‰, 0.05–0.24‰; Figure 3c). In Tibetan lakes, Yin, Feng, et al. (2016) observed highly negative δ^{202} Hg, similar to terrestrial sources (Demers et al., 2013; Jiskra et al., 2015, 2017), and positive Δ^{199} Hg in the sediments, which then increased over time (Figure 3d). The authors suggested that mercury introduced via watershed runoff was subjected to a higher degree of photo-reduction in the water column with declining ice cover. As such, the fact that the changes in local sources and/or processes may result in significant temporal trends in Δ^{199} Hg in certain lake sediments suggests that the positive Δ^{199} Hg shifts are unlikely to be a global phenomenon. Instead, the widespread and significant positive δ^{202} Hg shifts observed in the lake sediment cores indicate that δ^{202} Hg can serve as an important proxy for assessing temporal changes in anthropogenic mercury source contribution.

3.3. Importance of Baseline Mercury Isotope Ratios

The evaluation of the temporal trends in the lake sediment cores indicates that the increased contribution of local to regional anthropogenic mercury sources, which have a similar Δ^{199} Hg and a much higher δ^{202} Hg compared to the pre-industrial sediments, explain the differences in the temporal trends between δ^{202} Hg and Δ^{199} Hg. While the negative pre-industrial δ^{202} Hg resulted in the widespread positive shifts in δ^{202} Hg, the near-zero pre-industrial Δ^{199} Hg resulted in either absence or small positive shifts in Δ^{199} Hg in the lake sediment cores. To further evaluate the importance of pre-industrial mercury isotope ratios for explaining the temporal trends, we estimate the magnitude of δ^{202} Hg and Δ^{199} Hg changes from the pre-industrial to the present-day period except for Laguna Negrilla, which decreased in THg. In this assessment, the average sediment mercury isotope ratios reflecting the pre-industrial period are set as a baseline or a starting point of each core and the sediment δ^{202} Hg and Δ^{199} Hg for each dated year are subtracted from the average pre-industrial δ^{202} Hg and Δ^{199} Hg (baseline) to quantify the magnitude of changes in mercury isotope ratios over time. As illustrated in Figure 4, the lake sediment cores with the most negative δ^{202} Hg (-4.24 to -2.34‰) and Δ^{199} Hg baselines (-0.72 to -0.43‰) increased dramatically in δ^{202} Hg (by 0.82-1.95‰) and Δ^{199} Hg (by 0.20–0.70%) in the present-day period. The magnitude of δ^{202} Hg and Δ^{199} Hg increases from the baseline to the present-day period decreased as the baseline δ^{202} Hg and Δ^{199} Hg increased in each sediment core. Moreover, we found that the magnitude of δ^{202} Hg and Δ^{199} Hg changes did not correlate with the magnitude of THg changes (δ^{202} Hg; $r^2 = 0.09$, Δ^{199} Hg; $r^2 = 0.20$, both p > 0.05). Even among the unremote (δ^{202} Hg; $r^2 < 0.01$, Δ^{199} Hg; $r^2 = 0.17$, both p > 0.05) and remote sediments (δ^{202} Hg; $r^2 < 0.01$, Δ^{199} Hg; $r^2 = 0.07$, both p > 0.05), we did not observe significant relationships between the magnitude of δ^{202} Hg and Δ^{199} Hg changes and the magnitude of THg changes. This indicates that, unlike the THg concentration, the magnitude of mercury isotope ratio changes are not governed by the extent of mercury input.

Our results are in contrast with Lepak et al. (2020), who observed significant positive relationships between the magnitude of changes in mercury accumulation flux, which is correlated with the sediment THg



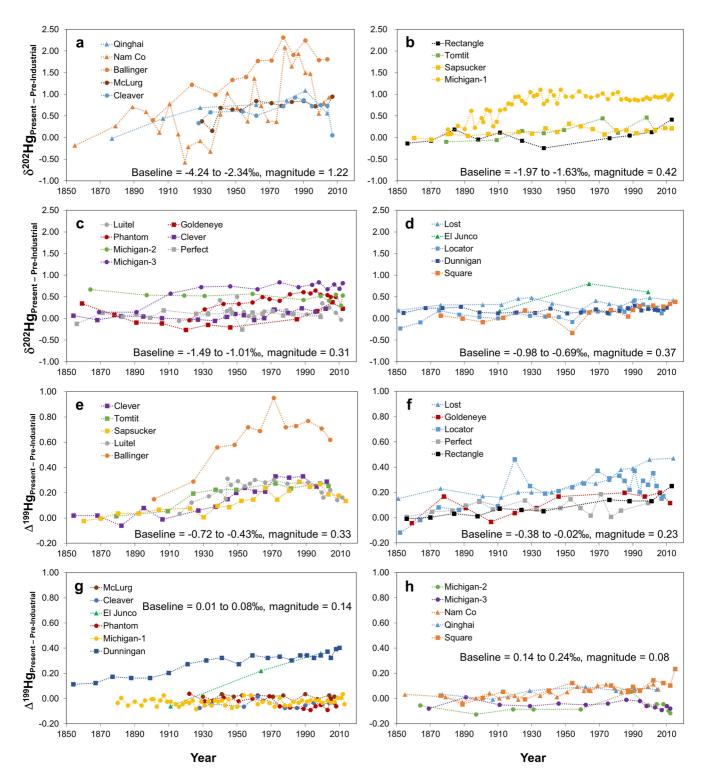


Figure 4. Magnitude of δ^{202} Hg (a to d) and Δ^{199} Hg changes (e to h) from the pre-industrial (1510–1850) period to the next core interval over time. From each panel, baseline represents the ranges of the pre-industrial δ^{202} Hg or Δ^{199} Hg. Magnitude indicates the average magnitude of δ^{202} Hg or Δ^{199} Hg changes from the pre-industrial to present-day period.

concentration, and the magnitude of changes in δ^{202} Hg and Δ^{199} Hg in nine remote lake sediments of North America. In that study, the authors suggested that the differences in the amount of atmospheric mercury input sourced from precipitation determine the inter-lake variability in the magnitude of mercury isotope ratio changes over time. Our compilation of 22 lake sediment cores suggests that the magnitude of changes in mercury isotope ratios from the pre-industrial to present-day period is governed by the pre-industrial variability in mercury isotope ratios. Even among the remote and unremote sediments, which showed drastic differences in their magnitude of THg change, the present-day mercury isotope ratios were statistically similar, supporting our conclusion that the changes in mercury isotope ratios in lake sediment cores reflect isotope mixing between anthropogenic mercury sources and pre-industrial sediments. Based on the isotope mixing effect, we expect that a further increase in the contribution of anthropogenic mercury sources to lake sediments would lead to cessation of further shifts in sediment mercury isotope ratios, while mercury concentration continues to increase via the input. Even within a single lake, Lake Michigan, Yin, Lepak, et al. (2016) observed differences in the pre-industrial sediment mercury isotope ratios across three sites $(\delta^{202}\text{Hg} = -1.63 \text{ to } -1.01\%, \Delta^{199}\text{Hg} = 0.06-0.20\%)$, and the magnitude of changes from the pre-industrial to present-day period (δ^{202} Hg; by 0.24–0.92, Δ^{199} Hg; by 0.01–0.06; Figures 1b and 1c) did not correlate with the magnitude of THg changes. This indicates that either within a single lake or across multiple lakes, the baseline δ^{202} Hg and Δ^{199} Hg plays a more important role in determining the magnitude of changes in mercury isotope ratios compared to THg.

3.4. Implication for the Minamata Convention on Mercury

The compilation of mercury isotope ratios in lake sediment cores suggests that δ^{202} Hg and Δ^{199} Hg of anthropogenic mercury sources and Δ^{200} Hg reflecting the extent of photo-oxidation are well preserved in lake sediments. In regards to temporal trends, we attribute the widespread positive shifts in δ^{202} Hg in the lake sediment cores (e.g., Guérdron et al., 2016; Lepak et al., 2020) to local to global rise in the modern-day anthropogenic mercury sources. The absence of widespread shifts in Δ^{199} Hg can be explained by a similar Δ^{199} Hg value between the pre-industrial sediments and modern-day anthropogenic mercury sources. The fact that the changes in local sources and/or processes may result in temporal trends in Δ^{199} Hg in certain lake sediments suggests that the positive Δ^{199} Hg shifts are unlikely to be a global phenomenon. While Δ^{199} Hg has been regarded as a conservative tracer for mercury sources due to limited processes resulting in MIF, our results suggest that the temporal evaluation of δ^{202} Hg, which has a much higher value compared to the pre-industrial sediments, is a more powerful proxy for evaluating long-term changes in anthropogenic mercury sources. Across a broad spatial scale, the characterization of pre-industrial mercury isotope ratios will be critical for deciphering between the absence and significant changes in mercury isotope ratios resulted from anthropogenic mercury source contribution.

Our study has wide-ranging implications for the global monitoring and effectiveness evaluation of the MC. Many previous studies have assessed temporal trends in mercury using atmospheric samples (Cheng et al., 2017; Dommergue et al., 2016) and biota (Blukacz-Richards et al., 2017; Lee et al., 2016; McKinney et al., 2017). While these samples provide insights into rapid responses to changes in mercury emissions and bioaccumulation, the isotope ratios reflecting mercury sources are susceptible to change via environmental, biological, and ecological conditions. For instance, previous studies have observed large seasonal and air mass differences in mercury isotope ratios in atmospheric samples (Fu et al., 2016; Huang et al., 2018). The tissues of biota including fish and birds have shown measurable MDF in the bioaccumulated mercury sources caused by metabolic processes (Kwon et al., 2013). The mercury isotope ratios of biota tissues also reflect a mixture of mercury sources, which may have been integrated over time and space (i.e., migratory sites) (Renedo et al., 2018, 2021). In fact, deciphering between policy signal and environmental noise has been one of the greatest challenges of evaluating the effectiveness of the MC. In this regard, the effective and long-term preservation of mercury isotope ratios of anthropogenic mercury sources in lake sediment cores can serve as an important proxy for assessing changes in anthropogenic mercury sources. The information on the pre-industrial mercury isotope ratios can also serve as an ideal and site-specific target for the effectiveness evaluation of the MC in which the large-scale reduction in local to regional anthropogenic mercury emissions would allow sediment mercury isotope ratios to recover to values similar to the pre-industrial sediment values. Lastly, when predicting future changes in mercury isotope ratios in sediment cores, the types of mercury emitting sectors governed under the MC and modifications in mercury

abatement technologies need to be considered. The intentional mercury use sectors including mining and manufacturing of commercial products, which predominantly emit Hg⁰, are subject to the most stringent global regulation (UNEP, 2019). Dramatic reductions in Hg⁰, which have the highest δ^{202} Hg relative to Hg²⁺ and Hg_p based on the anthropogenic mercury source end members estimated by Sun et al. (2016), may suppress further increases in δ^{202} Hg in sediment cores. The types of mercury abatement technologies in by-product sectors (i.e., fossil fuel combustion) are also known to determine δ^{202} Hg of the emitted mercury (Sun et al., 2014). We suggest that future studies that assess the timescale of mercury isotope turnover in relation to the reduction of anthropogenic mercury input and the sensitivity of local to global environmental and policy changes would enhance the utility of mercury isotopes measured in sediments for the monitoring and effectiveness evaluation of the MC.

Conflict of Interest

The authors declare no conflicts of interest.

Data Availability Statement

The compiled data used in this study is available at https://doi.org/10.5281/zenodo.5077887 (or https:// github.com/jhl329/lake-sediment-cores_compiled-data).

References

- Amos, H. M., Sonke, J. E., Obrist, D., Robins, N., Hagan, N., Horowitz, H. M., et al. (2015). Observational and constraints on global anthropogenic enrichment of mercury. *Environmental Science and Technology*, 49(7), 4036–4047. https://doi.org/10.1021/es5058665
- Appleby, P. G. (2002). Chronostratigraphic techniques in recent sediments. In *Tracking environmental change using lake sediments* (pp. 171–203): Springer.
 - Appleby, P. G., & Oldfield, F. (1978). The calculation of lead-210 dates assuming a constant rate of supply of unsupported 210Pb to the sediment. *Catena*, 5(1), 1–8. https://doi.org/10.1016/S0341-8162(78)80002-2
 - Appleby, P. G., & Oldfield, F. (1992). Applications of lead-210 to sedimentation studies. In Uranium-series disequilibrium: Applications to earth, marine, and environmental sciences. (2nd ed.). Clarendon Press.
 - Asmund, G., & Nielsen, S. P. (2000). Mercury in dated Greenland marine sediments. *The Science of the Total Environment*, 245(1–3), 61–72. https://doi.org/10.1016/S0048-9697(99)00433-7
 - Baskaran, M. (2016). Radon: A Tracer for Geological, Geophysical and Geochemical Studies (Vol. 367): Springer.
 - Benoit, J. M., Gilmour, C. C., Heyes, A., Mason, R. P., & Miller, C. L. (2003). Geochemical and biological controls over methylmercury production and degradation in aquatic ecosystems. ACS Symposium Series, 835, 262–297. https://doi.org/10.1021/bk-2003-0835.ch019 Bergquist, B. A., & Blum, J. D. (2007). Mass-dependent and -independent fractionation of Hg isotopes by photoreduction in aquatic sys-
 - tems. Science, 318(5849), 417–420. https://doi.org/10.1126/science.1148050 Bergquist, B. A., & Blum, J. D. (2009). The odds and evens of mercury isotopes: Applications of mass-dependent and mass-independent
 - isotope fractionation. *Elements*, 5(6), 353–357. https://doi.org/10.2113/gselements.5.6.353 Bessinger, B. A. (2014). Use of stable isotopes to identify sources of mercury in sediments: A review and uncertainty analysis. *Environmen*-
 - tal Forensics, 15(3), 265–280. https://doi.org/10.1080/15275922.2014.930939
 - Blaauw, M. (2010). Methods and code for 'classical'age- of radiocarbon sequences. *Quaternary Geochronology*, 5(5), 512–518. https://doi.org/10.1016/j.quageo.2010.01.002
 - Blukacz-Richards, E. A., Visha, A., Graham, M. L., McGoldrick, D. L., de Solla, S. R., Moore, D. J., & Arhonditsis, G. B. (2017). Mercury levels in herring gulls and fish: 42 years of spatio-temporal trends in the Great Lakes. *Chemosphere*, *172*, 476–487. https://doi.org/10.1016/j. chemosphere.2016.12.148
 - Blum, J. D., & Bergquist, B. A. (2007). Reporting of variations in the natural isotopic composition of mercury. Analytical and Bioanalytical Chemistry, 388, 353–359. https://doi.org/10.1007/s00216-007-1236-9

Blum, J. D., & Johnson, M. W. (2017). Recent developments in mercury stable isotope analysis. *Reviews in Mineralogy and Geochemistry*, 82(1), 733–757. https://doi.org/10.2138/rmg.2017.82.17

Blum, J. D., Sherman, L. S., & Johnson, M. W. (2014). Mercury isotopes in earth and environmental sciences. Annual Review of Earth and Planetary Sciences, 42, 249–269. https://doi.org/10.1146/annurev-earth-050212-124107

Cai, H., & Chen, J. (2016). Mass-independent fractionation of even mercury isotopes. *Science Bulletin*, 61(2), 116–124. https://doi.org/10.1007/s11434-015-0968-8

- Chen, J., Hintelmann, H., Feng, X., & Dimock, B. (2012). Unusual fractionation of both odd and even mercury isotopes in precipitation from Peterborough, ON, Canada. *Geochimica et Cosmochimica Acta*, 90, 33–46. https://doi.org/10.1016/j.gca.2012.05.005
- Cheng, I., Zhang, L., Castro, M., & Mao, H. (2017). Identifying Changes in Source Regions Impacting Speciated Atmospheric Mercury at a Rural Site in the Eastern United States. *Journal of the Atmospheric Sciences*, 74(9), 2937–2947. https://doi.org/10.1175/JAS-D-17-0086.1
- Cooke, C. A., Hintelmann, H., Ague, J. J., Burger, R., Biester, H., Sachs, J. P., & Engstrom, D. R. (2013). Use and legacy of mercury in the Andes. *Environmental Science and Technology*, 47(9), 4181–4188. https://doi.org/10.1021/es3048027
- Cooke, C. A., Martínez-Cortizas, A., Bindler, R., & Gustin, M. S. (2020). Environmental archives of atmospheric Hg deposition A review. *The Science of the Total Environment*, 709, 134800. https://doi.org/10.1016/j.scitotenv.2019.134800] 134800Refstyled

Demers, J. D., Blum, J. D., & Zak, D. R. (2013). Mercury isotopes in a forested ecosystem: Implications for air-surface exchange dynamics and the global mercury cycle. *Global Biogeochemical Cycles*, 27(1), 222–238. https://doi.org/10.1002/gbc.20021

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- Dommergue, A., Martinerie, P., Courteaud, J., Witrant, E., & Etheridge, D. M. (2016). A new reconstruction of atmospheric gaseous elemental mercury trend over the last 60 years from Greenland firn records. *Atmospheric Environment*, *136*, 156–164. https://doi. org/10.1016/j.atmosenv.2016.04.012
- Eakins, J. A., & Morrison, R. T. (1978). A new procedure for the determination of lead-210 in lake and marine sediments. *The International Journal of Applied Radiation and Isotopes*, 29(9–10), 531–536. https://doi.org/10.1016/0020-708X(78)90161-8
- Engstrom, D. R., Balogh, S. J., & Swain, E. B. (2007). History of mercury inputs to Minnesota lakes: Influences of watershed disturbance and localized atmospheric deposition. *Limnology & Oceanography*, 52(6), 2467–2483. https://doi.org/10.4319/lo.2007.52.6.2467
- Engstrom, D. R., Fitzgerald, W. F., Cooke, C. A., Lamborg, C. H., Drevnick, P. E., Swain, E. B., et al. (2014). Atmospheric Hg emissions from preindustrial gold and silver extraction in the Americas: A reevaluation from lake-sediment archives. *Environmental Science and Technology*, 48(12), 6533–6543. https://doi.org/10.1021/es405558e
- Enrico, M., Le Roux, G., Heimbürger, L. E., Van Beek, P., Souhaut, M., Chmeleff, J., & Sonke, J. E. (2017). Holocene atmospheric mercury levels reconstructed from peat bog mercury stable isotopes. *Environment Science and Technology*, 51(11), 5899–5906. https://doi. org/10.1021/acs.est.6b05804
- Estrade, N., Carignan, J., Sonke, J. E., & Donard, O. F. (2009). Mercury isotope fractionation during liquid-vapor evaporation experiments. Geochimica et Cosmochimica Acta, 73(10), 2693–2711. https://doi.org/10.1016/j.gca.2009.01.024
- Fitzgerald, W. F., Engstrom, D. R., Lamborg, C. H., Tseng, C. M., Balcom, P. H., & Hammerschmidt, C. R. (2005). Modern and historic atmospheric mercury fluxes in northern Alaska: Global sources and Arctic depletion. *Environmental Science and Technology*, 39, 557–568. https://doi.org/10.1021/es049128x
- Fu, X., Marusczak, N., Wang, X., Gheusi, F., & Sonke, J. E. (2016). Isotopic composition of gaseous elemental mercury in the free troposphere of the Pic du Midi Observatory. *France. Environmental Science and Technology*, 50(11), 5641–5650. https://doi.org/10.1021/acs. est.6b00033
- Fuller, C. C., van Geen, A., Baskaran, M., & Anima, R. (1999). Sediment chronology in San Francisco Bay, California, defined by 210Pb, 234Th, 137Cs, and 239,240 Pu. Marine Chemistry, 64(1–2), 7–27. https://doi.org/10.1016/S0304-4203(98)00081-4
- Ghosh, S., Schauble, E. A., Couloume, G. L., Blum, J. D., & Bergquist, B. A. (2013). Estimation of nuclear volume dependent fractionation of mercury isotopes in equilibrium liquid-vapor evaporation experiments. *Chemical Geology*, 336, 5–12. https://doi.org/10.1016/j. chemgeo.2012.01.008
- Goldberg, E. D. (1963). Geochronology with 210Pb radioactive dating (Vol. 121, p. 130). International Atomic Energy Agency
- Gray, J. E., Pribil, M. J., Van Metre, P. C., Borrok, D. M., & Thapalia, A. (2013). Identification of contamination in a lake sediment core using Hg and Pb isotopic compositions, Lake Ballinger, Washington, USA. Applied Geochemistry, 29, 1–12. https://doi.org/10.1016/j. apgeochem.2012.12.001
- Guédron, S., Amouroux, D., Sabatier, P., Desplanque, C., Develle, A. L., Barre, J., et al. (2016). A hundred year record of industrial and urban development in French Alps combining Hg accumulation rates and isotope composition in sediment archives from Lake Luitel. *Chemical Geology*, 431, 10–19. https://doi.org/10.1016/j.chemgeo.2016.03.016
- Huang, S., Sun, L., Zhou, T., Yuan, D., Du, B., & Sun, X. (2018). Natural stable isotopic compositions of mercury in aerosols and wet precipitations around a coal-fired power plant in Xiamen, southeast China. Atmospheric Environment, 173, 72–80. https://doi.org/10.1016/j. atmosenv.2017.11.003
- Jiskra, M., Wiederhold, J. G., Skyllberg, U., Kronberg, R. M., Hajdas, I., & Kretzschmar, R. (2015). Mercury deposition and reemission pathways in boreal forest soils investigated with Hg isotope signatures. *Environmental Science and Technology*, 49(12), 7188–7196. https:// doi.org/10.1021/acs.est.5b00742
- Jiskra, M., Wiederhold, J. G., Skyllberg, U., Kronberg, R. M., & Kretzschmar, R. (2017). Source tracing of natural organic matter bound mercury in boreal forest runoff with mercury stable isotopes. *Environmental Science: Processes and Impacts*, 19(10), 1235–1248. https:// doi.org/10.1039/C7EM00245A
- Klump, J. V., Edgington, D. N., Sager, P. E., & Robertson, D. M. (1997). Sedimentary phosphorus cycling and a phosphorus mass balance for the Green Bay (Lake Michigan) ecosystem. *Canadian Journal of Fisheries and Aquatic Sciences*, 54(1), 10–26. https://doi.org/10.1139/ f96-247
- Kurz, A. Y., Blum, J. D., Washburn, S. J., & Baskaran, M. (2019). Changes in the mercury isotopic composition of sediments from a remote alpine lake in Wyoming, USA. The Science of the Total Environment, 669, 973–982. https://doi.org/10.1016/j.scitotenv.2019.03.165
- Kwon, S. Y., Blum, J. D., Chirby, M. A., & Chesney, E. J. (2013). Application of mercury isotopes for tracing trophic transfer and internal distribution of mercury in marine fish feeding experiments. *Environmental Toxicology & Chemistry*, 32, 2322–2330. https://doi. org/10.1002/etc.2313
- Kwon, S. Y., Blum, J. D., Yin, R., Tsui, M. T. K., Yang, Y. H., & Choi, J. W. (2020). Mercury stable isotopes for monitoring the effectiveness of the Minamata Convention on Mercury. *Earth-Science Reviews*, 203, 103111. https://doi.org/10.1016/j.earscirev.2020.103111
- Lee, C. S., Lutcavage, M. E., Chandler, E., Madigan, D. J., Cerrato, R. M., & Fisher, N. S. (2016). Declining mercury concentrations in bluefin tuna reflect reduced emissions to the North Atlantic Ocean. *Environmental Science and Technology*, 50(23), 12825–12830. https://doi. org/10.1021/acs.est.6b04328
- Lepak, R. F., & Janssen, S. E. (2020). Mercury concentrations and isotopic compositions in sediment cores from North American lakes (Alaska, Minnesota, and Newfoundland). U.S. Geological Survey Data Release. https://doi.org/10.5066/P915RL9C
- Lepak, R. F., Janssen, S. E., Engstrom, D. R., Krabbenhoft, D. P., Tate, M. T., Yin, R., et al. (2020). Resolving atmospheric mercury loading and source trends from isotopic records of remote North American Lake Sediments. *Environmental Science and Technology*, 54(15), 9325–9333. https://doi.org/10.1021/acs.est.0c00579
- Li, C., Sonke, J. E., Le Roux, G., Piotrowska, N., Van der Putten, N., Roberts, S. J., et al. (2020). Unequal anthropogenic enrichment of mercury in Earth's northern and southern hemispheres. ACS Earth and Space Chemistry, 4(11), 2073–2081. https://doi.org/10.1021/ acsearthspacechem.0c00220
- Lindberg, S., Bullock, R., Ebinghaus, R., Engstrom, D., Feng, X., Fitzgerald, W., et al. (2007). A synthesis of progress and uncertainties in attributing the sources of mercury in deposition. *Ambio*, 19–32. https://doi.org/10.1579/0044-7447(2007)36[19:asopau]2.0.co;2. https:// www.jstor.org/stable/4315781
- Ma, J., Hintelmann, H., Kirk, J. L., & Muir, D. C. (2013). Mercury concentrations and mercury isotope composition in lake sediment cores from the vicinity of a metal smelting facility in Flin Flon, Manitoba. *Chemical Geology*, 336, 96–102. https://doi.org/10.1016/j. chemgeo.2012.10.037
- McKinney, M. A., Atwood, T. C., Pedro, S., & Peacock, E. (2017). Ecological change drives a decline in mercury concentrations in southern Beaufort Sea polar bears. *Environmental Science and Technology*, 51(14), 7814–7822. https://doi.org/10.1021/acs.est.7b00812



- Mergler, D., Anderson, H. A., Chan, L. H. M., Mahaffey, K. R., Murray, M., Sakamoto, M., & Stern, A. H. (2007). Methylmercury exposure and health effects in humans: A worldwide concern. AMBIO: A Journal of the Human Environment, 36(1), 3–11. https://doi.org/10.157 9/0044-7447(2007)36[3:MEAHEI]2.0.CO;2
- Mulvaney, K. M., Selin, N. E., Giang, A., Muntean, M., Li, C. T., Zhang, D., et al. (2020). Mercury benefits of climate policy in China: Addressing the Paris agreement and the Minamata convention simultaneously. *Environmental Science and Technology*, 54, 1326–1335. https://doi.org/10.1021/acs.est.9b06741
- Percival, J. B., & Outridge, P. M. (2013). A test of the stability of Cd, Cu, Hg, Pb and Zn profiles over two decades in lake sediments near the Flin Flon Smelter, Manitoba, Canada. *The Science of the Total Environment*, 454, 307–318. https://doi.org/10.1016/j.scitotenv.2013.03.011
- Ren, L., Arkin, P., Smith, T. M., & Shen, S. S. (2013). Global precipitation trends in 1900–2005 from a reconstruction and coupled model simulations. *Journal of Geophysical Research: Atmospheres*, 118(4), 1679–1689. https://doi.org/10.1002/jgrd.50212
- Renedo, M., Amouroux, D., Duval, B., Carravieri, A., Tessier, E., Barre, J., et al. (2018). Seabird tissues as efficient biomonitoring tools for Hg isotopic investigations: Implications of using blood and feathers from chicks and adults. *Environmental Science and Technology*, 52(7), 4227–4234. https://doi.org/10.1021/acs.est.8b00422
- Renedo, M., Pedrero, Z., Amouroux, D., Cherel, Y., & Bustamante, P. (2021). Mercury isotopes of key tissues document mercury metabolic processes in seabirds. *Chemosphere*, 263, 127777. https://doi.org/10.1016/j.chemosphere.2020.127777
- Rydberg, J., Gälman, V., Renberg, I., Bindler, R., Lambertsson, L., & Martínez-Cortizas, A. (2008). Assessing the stability of mercury and methylmercury in a varved lake sediment deposit. *Environmental Science and Technology*, 42(12), 4391–4396. https://doi.org/10.1021/ es7031955
- Selin, N. E. (2009). Global biogeochemical cycling of mercury: A review. Annual Review of Environment and Resources, 34, 43–63. https:// doi.org/10.1146/annurev.environ.051308.084314
- Song, W., Li, A., Ford, J. C., Sturchio, N. C., Rockne, K. J., Buckley, D. R., & Mills, W. J. (2005). Polybrominated diphenyl ethers in the sediments of the Great Lakes. 2. Lakes Michigan and Huron. *Environmental Science and Technology*, 39(10), 3474–3479. https://doi. org/10.1021/es048291p
- Streets, D. G., Horowitz, H. M., Jacob, D. J., Lu, Z., Levin, L., Ter Schure, A. F., & Sunderland, E. M. (2017). Total mercury released to the environment by human activities. *Environmental Science and Technology*, 51(11), 5969–5977. https://doi.org/10.1021/acs.est.7b00451
- Sun, R., Sonke, J. E., Heimbürger, L. E., Belkin, H. E., Liu, G., Shome, D., et al. (2014). Mercury stable isotope signatures of world coal deposits and historical coal combustion emissions. *Environmental Science and Technology*, 48(13), 7660–7668. https://doi.org/10.1021/ es501208a
- Sun, R., Streets, D. G., Horowitz, H. M., Amos, H. M., Liu, G., Perrot, V., et al. (2016). Historical (1850–2010) mercury stable isotope inventory from anthropogenic sources to the atmosphere. Mercury isotope emission inventory. *Elementa: Science of the Anthropocene*, 4. https://doi.org/10.12952/journal.elementa.000091
- Sunderland, E. M. (2007). Mercury exposure from domestic and imported estuarine and marine fish in the US seafood market. Environmental Health Perspectives, 115(2), 235–242. https://doi.org/10.1289/ehp.9377
- United Nations Environment Programme (UNEP). (2019). Minamata convention on mercury Text and annexes. Retrieved from http:// www.mercuryconvention.org/Portals/11/documents/Booklets/COP3-version/Minamata-Convention-booklet-Sep2019-EN.pdf
- Wasik, J. K. C., Engstrom, D. R., Mitchell, C. P. J., Swain, E. B., Monson, B. A., Balogh, S. J., et al. (2015). The effects of hydrologic fluctuation and sulfate regeneration on mercury cycling in an experimental peatland. *Journal of Geophysical Research: Biogeosciences*, 120(9), 1697–1715. https://doi.org/10.1002/2015JG002993
- Wiederhold, J. G., Cramer, C. J., Daniel, K., Infante, I., Bourdon, B., & Kretzschmar, R. (2010). Equilibrium mercury isotope fractionation between dissolved Hg (II) species and thiol-bound Hg. *Environmental Science and Technology*, 44(11), 4191–4197. https://doi. org/10.1021/es100205t
- Yin, R., Feng, X., Hurley, J. P., Krabbenhoft, D. P., Lepak, R. F., Kang, S., et al. (2016). Historical records of mercury stable isotopes in sediments of Tibetan lakes. *Scientific Reports*, 6: 23332. https://doi.org/10.1038/srep23332
- Yin, R., Lepak, R. F., Krabbenhoft, D. P., & Hurley, J. P. (2016). Sedimentary records of mercury stable isotopes in Lake Michigan. Elementa: Science of the Anthropocene, 4. https://doi.org/10.12952/journal.elementa.000086
- Zdanowicz, C. M., Krümmel, E. M., Poulain, A. J., Yumvihoze, E., Chen, J., Štrok, M., et al. (2016). Historical variations of mercury stable isotope ratios in Arctic glacier firm and ice cores. *Global Biogeochemical Cycles*, 30(9), 1324–1347. https://doi.org/10.1002/2016GB005411
- Zerkle, A. L., Yin, R., Chen, C., Li, X., Izon, G. J., & Grasby, S. E. (2020). Anomalous fractionation of mercury isotopes in the Late Archean atmosphere. *Nature Communications*, 11(1), 1–9. https://doi.org/10.1038/s41467-020-15495-3
- Zhou, J., Obrist, D., Dastoor, A., Jiskra, M., & Ryjkov, A. (2021). Vegetation uptake of mercury and impacts on global cycling. *Nature Reviews Earth & Environment*, 2, 1–284. https://doi.org/10.1038/s43017-021-00146-y